

2 Using the multivariate Hawkes process to study interactions between multiple
3 species from camera trap data

4 Lisa Nicvert¹, Sophie Donnet², Mark Keith³, Mike Peel^{4, 5, 6}, Michael J. Somers³, Lourens H.
5 Swanepoel⁷, Jan Venter^{8, 9}, Hervé Fritz^{9, 10}, and Stéphane Dray¹

6 ¹Université de Lyon, Université Lyon 1, CNRS, VetAgro Sup, Laboratoire de Biométrie et Biologie Evolutive, UMR5558, Villeurbanne, France

7 ²Paris-Saclay University, AgroParisTech, INRAE, UMR MIA-Paris, France

8 ³Eugène Marais Chair of Wildlife Management, Mammal Research Institute, Department of Zoology and Entomology, University of Pretoria,
9 Pretoria, South Africa

10 ⁴Agricultural Research Council, Animal Production Institute, Rangeland Ecology, Pretoria, South Africa

11 ⁵School of Animal, Plant and Environmental Sciences, University of the Witwatersrand, Johannesburg, South Africa

12 ⁶Applied Behavioural Ecology and Ecosystems Research Unit, University of South Africa, Florida, South Africa

13 ⁷Department of Biological Sciences, Faculty of Science, Engineering and Agriculture, University of Venda, Thohoyandou, 0950, South Africa

14 ⁸Department of Conservation Management, Faculty of Science, George Campus, Nelson Mandela University, George, South Africa

15 ⁹REHABS, International Research Laboratory, CNRS-NMU-UCBL, Nelson Mandela University, George, South Africa

16 ¹⁰Sustainability Research Unit, Nelson Mandela University, George, South Africa

17 Correspondence: Lisa Nicvert (lisa.nicvert@univ-lyon1.fr)

18 Authorship statement: Hervé Fritz and Stéphane Dray are co-last authors.

19 Open Research: Data and code (Nicvert et al., 2023) are available on Figshare at

20 <https://doi.org/10.6084/m9.figshare.24552157.v3>.

21 **Keywords**— African savanna; camera trap; interaction network; inter-specific interactions; multivariate
22 Hawkes process; reactive response; Snapshot Safari; spatio-temporal interactions

Abstract

Inter-specific interactions can influence species' activity and movement patterns. In particular, species may avoid or attract each other through reactive responses in space and/or time. However, data and methods to study such reactive interactions have remained scarce and generally limited to two interacting species. Nowadays, the deployment of camera traps opens new opportunities but adapted statistical techniques are still required to analyze interaction patterns with such data. We present the multivariate Hawkes process (MHP) and show how it can be used to analyze interactions between several species using camera trap data. Hawkes processes use flexible pairwise interaction functions, allowing us to consider asymmetries and variations over time when depicting reactive temporal interactions. After describing the theoretical foundations of the MHP, we outline how its framework can be used to study inter-specific interactions with camera trap data. We design a simulation study to evaluate the performance of the MHP and of another existing method to infer interactions from camera trap-like data. We also use the MHP to infer reactive interactions from real camera trap data for five species from South African savannas (impala *Aepyceros melampus*, greater kudu *Tragelaphus strepsiceros*, lion *Panthera leo*, blue wildebeest *Connochaetes taurinus* and Burchell's zebra *Equus quagga burchelli*). The simulation study shows that the MHP can be used as a tool to benchmark other methods of inter-specific interactions inference and that this model can reliably infer interactions when enough data is considered. The analysis of real data highlights evidence of predator avoidance by prey and herbivore-herbivore attraction. Lastly, we present the advantages and limits of the MHP and discuss how it can be improved to infer attraction/avoidance patterns more reliably. As camera traps are increasingly used, the multivariate Hawkes process provides a promising framework to decipher the complexity of interactions structuring ecological communities.

1 Introduction

Inter-specific interactions affect many aspects of ecological communities. For instance, they influence ecosystem services (Valiente-Banuet et al., 2015), species assembly via biotic filtering (Ovaskainen et al., 2017) and the behavior of interacting species. In particular, interactions are one of the factors that structure the way in which animal species move in the landscape and adjust their habitat choices or activity times (Palmer et al., 2022). Mobile animals can respond to interactions by avoiding or seeking proximity with individuals of other species, depending on the positive or negative outcome of the interactions. For instance, prey can avoid their predators (Say-Sallaz et al., 2019), competing species can avoid each other (Cornhill et al., 2022; Searle et al., 2021), or herbivores can forage together to reduce predation risk or increase access to preferred foraging resources (Beaudrot et al., 2020). In this paper, we will use the term “interaction” to refer to the attraction or avoidance of a species by another one, even though “interaction” also refers to the underlying process to the attraction/avoidance pattern.

These interactions (as defined above) can occur in space and/or time, at different scales. Species can adjust their space use in response to the expected distribution of other species (proactive spatial interaction, Palmer et al., 2022). Species can also alter their daily activity patterns (e.g. Karanth et al., 2017) in response to other species (proactive temporal interaction). However, some species could also exhibit a reactive response to the presence of other species, i.e. change their behavior in response to the actual presence of a species sometime before at a given location (e.g. Karanth et al., 2017; Parsons et al., 2016). This type of response could be mediated, for instance, by olfactory (Cornhill and Kerley, 2020; Kuijper et al., 2014) or auditory cues (Hettena et al., 2014). Investigating these reactive interactions is particularly promising as it allows us to identify fine-grained patterns that could be missed by approaches aggregating data in space or in time (Cusack et al., 2017; Frey et al., 2017; Parsons et al., 2022).

Investigating such fine-scale responses is very challenging, as it requires an intensive sampling effort to monitor multiple species in space and time. In this context, camera traps open new opportunities to study the spatial and temporal activities of multiple species (Caravaggi et al., 2017). Camera trap arrays allow for

72 the collection of multiple species occurrences and, therefore, the monitoring of entire communities for large
73 areas continuously in time (Pardo et al., 2021). Hence, camera traps can produce massive amounts of data
74 and offer new possibilities to study interactions between several species at multiple scales. Moreover, they
75 are relatively cheap and easier to set up than classical fieldwork survey techniques (e.g. transects),
76 especially for rare or elusive species or in remote areas. As camera traps become more affordable and
77 automated species identification methods from pictures are being developed with deep learning, camera
78 trap data (and other passive sensors data) will likely become more abundant in the future (Caravaggi et al.,
79 2017).

80 With camera trap data, interspecific interactions are mostly studied at a broad spatial or temporal scale. To
81 do this, data are often aggregated so that either the spatial or the temporal aspect is completely ignored.
82 There are two main approaches for this purpose: comparing species' daily activities patterns (Ridout and
83 Linkie, 2009) or spatial occupancy patterns (e.g. with the multispecies occupancy model of Rota et al.,
84 2016). Such methods provide a measure of the proactive attraction or avoidance strategy, with species
85 adapting their space or time use in anticipation to other species' presence or absence (Palmer et al., 2022).
86 However, other approaches have combined spatial and temporal aspects to infer reactive
87 attraction/avoidance strategies (frequently called *spatio-temporal interactions* in the literature (Karanth
88 et al., 2017; Murphy et al., 2021; Niedballa et al., 2019; Prat-Guitart et al., 2020)). Most methods using
89 camera trap data quantify only the temporal aspect of reactive interactions: therefore, we will call the
90 inferred patterns *reactive temporal interactions*. The majority are based on the computation of time
91 intervals between the detections of two species at a given place (e.g., Harmsen et al., 2009): here, we call
92 this family of methods *inter-event times methods*. The distribution of time intervals can then be contrasted
93 according to the order of appearance of species (Parsons et al., 2016; Prat-Guitart et al., 2020) or
94 summarized by a statistic that is compared to values obtained under a null model (usually data permutation,
95 Cusack et al., 2017; Galindo-Aguilar et al., 2022; Karanth et al., 2017; Murphy et al., 2021). For a
96 comparison of different approaches to infer reactive temporal avoidance with time interval measures, see

97 [Niedballa et al. \(2019\)](#). Other more recent approaches use point processes, which allow us to integrate
98 temporal dependence in a model-based framework ([Kellner et al., 2022](#); [Schliep et al., 2018](#)).

99 Although all methods described above are useful to study reactive interactions, they usually focus on pairs
100 of species and can therefore be unsuitable for studying complex interaction networks. For instance, these
101 methods can identify spurious interactions between two species if other species are involved in the
102 interaction network but not considered in the analysis. Moreover, they summarize the effect of a species on
103 another one by a single value (e.g., the median of the time interval, [Karanth et al., 2017](#)), thus ignoring the
104 multiscale and possibly time-dependent changes in the attraction/avoidance patterns (but see [Cusack et al.,](#)
105 [2017](#)).

106 In this paper, we propose the multivariate Hawkes process (MHP) ([Hawkes, 1971](#); [Lambert et al., 2018](#)) as
107 a modeling framework to infer reactive interactions between multiple species from passive sensors such as
108 camera traps. Hawkes processes belong to the family of point processes which allow for the analysis of
109 species capture events in continuous time, thus avoiding any data aggregation procedure. In Hawkes
110 processes, species' interactions are modeled as pairwise interaction functions which depend on the time
111 elapsed between species detections. The MHP used in this article is generative and offers the possibility to
112 simulate occurrence data, given parameters specification. It also comes with an inference procedure which
113 allows to adjust the pairwise interaction functions from observed data. It deals properly with indirect effects
114 caused by species interaction chains, thus minimizing the risk of inferring spurious interactions. We believe
115 that this model is a useful conceptual framework which is well-suited for assessing reactive temporal
116 interactions between species. We first present the Hawkes process and how it can be used to analyze camera
117 trap data. Then, we describe the MHP used in this article, which was developed by [Lambert et al. \(2018\)](#)
118 and implemented in the R package `UnitEvents` ([Albert et al., 2021](#)). We then show how this model can be
119 used to simulate data to evaluate the performance of statistical methods or to infer interactions from camera
120 trap data. We also apply the MHP on real camera trap data from the Snapshot Safari monitoring program
121 ([Pardo et al., 2021](#)) to infer reactive temporal interactions between five mammal species. Finally, we

122 discuss on the usefulness of the MHP and the perspectives on how to develop this model further.

123 **2 Material and methods**

124 All analyses were performed using R statistical software (v4.3.0; [R Core Team, 2023](#)) and the code and data
125 ([Nicvert et al., 2023](#)) are available at <https://doi.org/10.6084/m9.figshare.24552157.v3>.

126 **2.1 Model: the multivariate Hawkes process (MHP)**

127 Hawkes processes are a family of point processes used to describe dependencies between punctual events.

128 These processes belong to the class of self-exciting point processes for which the probability of occurrence

129 at time t depends on the previous events occurrences. The first Hawkes process was introduced in 1971 by

130 Alan G. Hawkes ([Hawkes, 1971](#)). Originally applied to model aftershocks following earthquakes (e.g.

131 [Ogata, 1988](#)), Hawkes processes have been applied in various fields ([Reinhart, 2018](#)), for instance to model

132 crime recurrence in cities ([Mohler et al., 2018](#)), the evolution of prices on the stock market ([Hawkes, 2018](#))

133 or the transmission of action potentials in a network of neurons ([Reynaud-Bouret et al., 2013](#)). Theoretical

134 properties of Hawkes processes have also been thoroughly studied, and numerous extensions have been

135 proposed.

136 Throughout this article, we define an occurrence as a detection of an individual at a camera at a given time,

137 and we do not take imperfect detection into account. To describe the model, we consider data on the

138 occurrences of S species collected on C cameras. In our framework, the data collected on C cameras are

139 seen as C independent realizations of the MHP. Let T_m^{li} denote the m -th instant of punctual occurrence for

140 species i at camera l . Let N_i^l be the total number of occurrences for species i at camera l . We model the

141 occurrence times $(T_m^{li})_{m=1\dots N_i^l, i=1\dots S, l=1\dots C}$ as C realizations of a MHP.

142 To model punctual occurrences, point processes use a latent intensity function, which is a measure of the

143 rate at which events occur in time. When modeling species occurrences from camera trap data, the intensity

144 for a given species represents the rate at which this species occurs at a camera. For species i , the intensity

145 $\lambda_i^l(t)$ at camera l is formally defined as (Daley and Vere-Jones, 2003):

$$\lambda_i^l(t) = \lim_{\delta \rightarrow 0} \frac{\mathbb{P}\{n_{[t, t+\delta]}^{li} > 0\}}{\delta} \quad (1)$$

146 where $n_{[t, t+\delta]}^{li}$ is the number of points occurring between times t and $t + \delta$ for species i at camera l and δ is
 147 an infinitesimally small amount of time. Informally, the intensity of a point process multiplied by a small
 148 amount of time can be viewed as the probability that there will be at least one point occurring around time t .
 149 In this article, we use the R package `UnitEvents` (Albert et al., 2021), available at
 150 https://sourcesup.renater.fr/frs/?group_id=3267, to simulate and infer MHPs. `UnitEvents` is only available
 151 on Linux and Mac OS. However, in the code and data repository for the article (Nicvert et al., 2023), we
 152 provide a Dockerfile allowing to run the analyses from any operating system (including Windows).
 153 `UnitEvents` implements the MHP described in Lambert et al. (2018). In this framework, the intensity of
 154 species i seen on camera l for a Hawkes process with S interacting species is written as:

$$\lambda_i^l(t) = \left(\nu_i + \sum_{j=1}^S \sum_{m \mid T_m^{lj} < t} f_{j \rightarrow i}(t - T_m^{lj}) \right)_+ \quad (2)$$

155 where $\lambda_i^l(t)$ represents the intensity for species i (as defined above) at camera l . ν_i is a positive parameter,
 156 the *background rate*: it represents the basal intensity of species i (in time^{-1} , e.g., day^{-1}) unrelated to
 157 previous occurrences. For instance, ν_i would be low for a rare species and higher for a common species.
 158 $f_{j \rightarrow i}$ is the *interaction function* which represents the influence of an occurrence of species j on species i as
 159 a function of time delay: positive values of $f_{j \rightarrow i}$ represent an attraction of species i by species j , negative
 160 values represent a repulsion and null values independence. In the case $j = i$, the function $f_{i \rightarrow i}$ represents
 161 the interaction between individuals of the same species i . In that case, we will call $f_{i \rightarrow i}$ the *auto-interaction*
 162 *function*: it could reflect for instance the fact that some species are solitary or gregarious. $f_{j \rightarrow i}$ are defined
 163 as piecewise constant functions with K time bins of equal length δ :

$$f_{j \rightarrow i} = \sum_{k=1}^K a_{j \rightarrow i}^k \mathbf{1}_{(k-1)\delta, k\delta]} \quad (3)$$

164 where $\mathbf{1}_{(k-1)\delta, k\delta]}$ denotes the indicator function between delays $(k-1)\delta$ and $k\delta$. The K coefficients $a_{j \rightarrow i}^k$
 165 represent the average number of occurrences of species i gained (if positive) or suppressed (if negative) by
 166 an occurrence of species j in the k -th interval after this occurrence of species j .

167 In this framework, $f_{j \rightarrow i}$ can take negative values, thus allowing to model repulsive effect of species j on
 168 species i . As the intensity λ_i^l must be positive by definition, Equation (2) includes a positive part $(\cdot)_+$.
 169 However, for mathematical reasons, in the following developments we will assume that the negative values
 170 of $f_{j \rightarrow i}$ are never too strong so that the intensity never becomes negative, and the positive part is not
 171 needed. To enforce this assumption, the repulsion terms can only be as strong as the other terms making up
 172 the total intensity.

173 Figure 1 illustrates a realization of a MHP with five species (measured at a single camera) simulated with
 174 `UnitEvents`. In this example, some species attract each other (see the interaction network on Figure 1a)
 175 with the same decreasing discrete exponential interaction function with $K = 12$ time bins of width $\delta = 4$
 176 hours (Figure 1b). The background rate is the same for all species and is fixed to 0.2 occurrences day^{-1} .
 177 The right panel (Figure 1c) shows the simulated species occurrences and associated intensities over time.
 178 When nothing happens, the intensity is fixed at the background rate. When an attracting species occurs, the
 179 intensity of the attracted species peaks, making an occurrence more likely. For instance, each occurrence of
 180 species s_1 gives rise to a peak in the intensity of s_2 . Moreover, when several attracting events occur, the
 181 interaction functions add up, which makes the occurrence of the target species even more likely.

182 2.2 Model inference

183 The inference procedure implemented in the `UnitEvents` package is a fast and scalable LASSO-penalized
 184 (least absolute shrinkage and selection operator) least-squares criterion. It allows to estimate a single MHP
 185 from C realizations (in our setting, this corresponds to C cameras).

186 Let $(\beta_1, \dots, \beta_S)$ denote the parameters of interest for each species $i=1, \dots, S$. Each β_i is a vector of size
 187 $1 + SK$ containing the background rate of species i (ν_i) and the parameters of the interaction functions
 188 targeted to this species i for the S species and the K bins: $\beta_i = (\nu_i, (a_{j \rightarrow i}^k)_{j=1 \dots S, k=1 \dots K})$. Each β_i is
 189 estimated as:

$$\widehat{\beta}_i = \arg \min_{\beta_i} \text{LASSO}(\beta_i) \quad \text{where} \quad \text{LASSO}(\beta_i) = \underbrace{-2 \sum_{l=1}^C \mathbf{b}_i^{lT} \beta_i + \beta_i^T \sum_{l=1}^C \mathbf{G}^l \beta_i}_{\text{least-squares}} + \underbrace{2 \mathbf{d}_i^T |\beta_i|}_{\text{penalization}} \quad (4)$$

190 where T denotes transposition and $|\beta_i|$ is the vector containing the absolute values of the coordinates of β_i .
 191 \mathbf{b}_i^l is an observable vector of size $1 + SK$. If camera l is active between times α_l and η_l , then

$$\mathbf{b}_i^l = \left(N_i^l, \left(\int_{\alpha_l}^{\eta_l} n_{[t-k\delta, t-(k-1)\delta]}^{lj} \, dn_t^{li} \right)_{j=1 \dots S, k=1 \dots K} \right). \quad (5)$$

192 Its first value is the total count of species i observed on camera l . The other values represent the total
 193 occurrence counts of the species j observed in the k -th bin before the occurrences of species i at camera l .
 194 \mathbf{G}^l is also an observable matrix defined as:

$$\mathbf{G}^l = \int_{\alpha_l}^{\eta_l} \mathbf{c}_t^l \mathbf{c}_t^{lT} \, dt \quad (6)$$

195 where \mathbf{c}_t^l is a vector of size $1 + SK$ defined as $\mathbf{c}_t^l = \left(1, \left(n_{[t-k\delta, t-(k-1)\delta]}^{lj} \right)_{j=1 \dots S, k=1 \dots K} \right)$. Its first value
 196 is 1 and other values represent the occurrence counts of species j occurring on camera l in the k -th bin
 197 before time t .

198 The term $2\mathbf{d}_i^T |\beta_i|$ of Equation (4) corresponds to the LASSO penalization: it can make some parameter
 199 values shrink to zero and thus avoid overparameterization. The strength of this LASSO penalization is
 200 controlled by the weights vectors \mathbf{d}_i , which are computed from the data and tuned by a unique user-chosen
 201 parameter γ (equation derived from Lambert et al. (2018) adapted from Hansen et al. (2015)):

$$\mathbf{d}_i = \sqrt{2\gamma \log(S + S^2K) \sum_{l=1}^C \int_{\alpha_l}^{\eta_l} \mathbf{c}_t^{l^2} dn_t^{li} + \frac{\gamma \log(S + S^2K)}{3} \max_{l=1 \dots C} \left(\sup_{t \in [\alpha_l, \eta_l]} |\mathbf{c}_t^l| \right)}. \quad (7)$$

202 The choice of a suitable value for γ is crucial for model selection, because γ ensures that only relevant
 203 nonzero parameters are kept in the model. However, choosing a good value for γ is difficult: it has been
 204 evaluated by simulations in Lambert et al. (2018) and Hansen et al. (2015), and we proceeded likewise in
 205 this article.

206 In the current implementation of `UnitEvents`, three flavors of the LASSO penalization are available. We
 207 choose the "Bernstein Vanishing LASSO" (BVL), where the penalization in Equation (4) is first applied to
 208 discard weak interaction parameters. Then, the estimates of the remaining non-null parameters are obtained
 209 by minimizing the least-squares criterion. Lastly, an additional step is introduced to remove parameters
 210 smaller than a data-computed threshold (see Lambert et al., 2018, for details and justification).

211 In the implementation of `UnitEvents`, the bins width δ and the number of bins K for the interaction
 212 functions are fixed by the user, who also needs to choose a value of γ a priori. The other parameters
 213 (interaction functions coefficients $a_{j \rightarrow i}^k$ and background rates ν_i) are fitted as described before.

214 **2.3 Simulation study**

215 We generated camera trap-like data under the MHP and used these simulated data to (i) evaluate the
 216 performance of a method and (ii) tune the penalization parameter for inference on real data.

217 **2.3.1 Simulation parameters**

218 For these two objectives, we conducted two sets of simulations in the same conditions. We considered an
 219 interaction network with five species $s_{i=1 \dots 5}$ where s_1 attracts s_2 and s_2 attracts s_3 and s_4 (network from
 220 Figure 1a). This network represents a difficult case as an inference method should detect direct interactions,
 221 but not spurious indirect interactions (e.g., $s_1 \rightarrow s_3$) and identify that species s_5 is not interacting with
 222 others. In this simulation, we define the true interactions by decreasing exponential functions until two days:

$$f(t) = \begin{cases} a \exp\left(-\frac{\ln(2)}{0.5}t\right) & \text{if } t < 2 \\ 0 & \text{if } t \geq 2 \end{cases} \quad (8)$$

223 where a is the interaction strength. The half-life of this function is the denominator of the decrease rate, so
 224 that this function will reach half of its initial value at $t = 0.5$ day. The interaction strength a for the true
 225 model varied from 0.01 to 1 day⁻¹. Here, the interaction strength represents the maximum intensity of the
 226 pairwise interaction function for $t = 0$. An analogous interaction function is shown in Figure 1b with $a = 1$
 227 and with discrete bins. The background rate was fixed at 0.1 day⁻¹ for all species.

228 The simulated trapping length varied from 20 to 500 trapping days for each camera over 25 cameras
 229 (making up to 12 500 trapping days in total). For each condition, 30 different data sets were generated to
 230 evaluate the variability of the inference.

231 We evaluated the performance of the inference by computing the true positive and true negative rates. The
 232 true positive rate is the proportion of inferred nonzero interactions over the count of true nonzero
 233 interactions. The true negative rate is the proportion of inferred null interactions over the count of true null
 234 interactions.

235 **2.3.2 Evaluating a method to infer reactive temporal interactions**

236 We illustrated how synthetic data generated with the MHP can be used to evaluate the performance of a
 237 method to infer inter-specific interactions, considering the inter-event times method of [Murphy et al.](#)
 238 (2021). We applied the method of [Murphy et al. \(2021\)](#) on simulated data (simulation settings are
 239 described in Section 2.3.1). This method consists in computing the median time between directed pairwise
 240 species occurrences (excluding pairs from the same species) for observed and randomly permuted data (999
 241 permutations). The permutation procedure involved randomly changing the cameras' labels of species
 242 occurrences (for details, see [Murphy et al., 2021](#)). Finally, the statistical significance of interactions was
 243 estimated by comparing the median time for observed and permuted data. We used a significance threshold
 244 of 5 % with a Holm correction for multiple testing.

245 **2.3.3 Choice of the penalization parameter**

246 To choose the best penalization parameter γ in the context of interactions inference, we used a simulation
247 approach (simulation settings are described in Section 2.3.1). We inferred MHPs with different values of γ
248 (between 0.3 and 1) from the simulated datasets. For the inference parameters, we chose $K = 12$ bins of
249 width $\delta = 4$ hours (2 days in total, corresponding to the length of the simulated interactions functions).
250 Then, we defined any inferred interaction function as null if all bins were zero over the function's support,
251 and non-null if at least one bin was not null.

252 **2.4 Application: analysis of interactions between five species in the African** 253 **savanna**

254 We used the MHP to infer interaction functions between five species of the southern African savanna:
255 impala *Aepyceros melampus*, greater kudu *Tragelaphus strepsiceros*, lion *Panthera leo*, blue wildebeest
256 *Connochaetes taurinus* and Burchell's zebra *Equus quagga burchelli*.

257 **2.4.1 Data collection**

258 Camera trap data were collected as part of the long-term Snapshot Safari monitoring program (Pardo et al.,
259 2021). Snapshot Safari is a network of camera trap grids set up in more than 30 locations in southern
260 Africa. The camera trap design consists in grids of 5 km² in each location, in which cameras were fixed at
261 about 50 cm high. Cameras were automatically triggered by motion or heat using passive infrared sensors.
262 Each camera was programmed to take a series of three images within 1–5 seconds of each other by day, and
263 only one image by night to minimize disturbance occasioned by white flash. For this analysis, we focused
264 on six camera trap grids in the savanna biome in northern South Africa: the Associated Private Nature
265 Reserves (around Kruger National Park), Kruger National Park, Madikwe Game Reserve, Pilanesberg
266 National Park, Somkhanda Game Reserve and Venetia Limpopo Nature Reserve (see Figure 2).

267 **2.4.2 Data pre-processing**

268 Pictures were classified by citizen science using the Zooniverse platform (www.zooniverse.org), where
269 pictures were available online and annotated by more than 150 000 volunteers (see [Pardo et al., 2021](#), for
270 more details).

271 For this analysis, we filtered out cameras where capture events were too rare (less than 2 pictures in total or
272 less than 1 picture every 30 days on average). We did not filter for independence between occurrences of
273 the same species. However, since the Hawkes model does not allow two capture events to occur
274 simultaneously, if two or more individuals of different species were seen on the same capture event, their
275 occurrence time was randomly shifted from one minute in advance to one minute later. For multiple
276 individuals of the same species seen simultaneously, the occurrences of the individuals were counted as a
277 single event (i.e. an occurrence corresponds to an individual or a group of individuals of a given species).
278 After the filtering procedure, there were 72 703 occurrence events (corresponding to 70 409 unique
279 pictures) collected on 179 cameras in total. Cameras were active during 503 ± 224 (sd) days on average
280 (min: 19 days, max: 851 days), amounting to 90 176 trapping days on all cameras. All pictures were taken
281 between June 2017 and November 2019.

282 **2.4.3 Parameters inference**

283 We inferred the parameters of a MHP using interaction functions defined by $K = 6$ bins of $\delta = 6$ hours (36
284 hours in total). This parametrization should allow us to capture the dynamics of reactive temporal
285 interactions with enough granularity while keeping a relatively low number of parameters to estimate to
286 allow reliable inference. Using results of the simulation study (see Section [3.1.2](#)), we decided to set the
287 value of the penalization parameter γ to 0.5.

3 Results

3.1 Simulation study

3.1.1 Evaluating a method to infer reactive temporal interactions

We used data simulated under the MHP to evaluate the method of [Murphy et al. \(2021\)](#). As expected, the ability to detect interactions (true positive rate) increases with the strength of the interactions (Figure 3). Provided the interaction strength a is big enough (at least 0.1 day^{-1}), the ability to detect interactions increases with the number of trapping days, which indicates that a significant sampling effort is required to infer interactions from camera trap data (at least 300 trapping days for 25 cameras when the interaction strength is above 0.2 day^{-1}). More surprisingly, when the interaction strength is high (at least 0.5 day^{-1}), the true negative rate decreases with increasing sampling effort. This indicates that the method wrongly detects interactions between non-interacting species. Additional investigations (Appendix S1: [Section S1](#)) show that these errors mainly concern the detection of spurious indirect interactions between species involved in interaction chains (e.g., $s_1 \rightarrow s_3$).

3.1.2 Choice of the penalization parameter

The simulation study to find suitable values for the penalization parameter γ led to the results shown in Figure 4. Unsurprisingly, the ability to detect true interactions (true positive rate) increases with the number of trapping days and the strength of interactions. When the penalization is too low ($\gamma = 0.3$; top row), the model tends to identify interactions between non-interacting species (reducing the true negative rate) but this problem vanishes when sampling effort increases. On the other hand, a high penalization ($\gamma = 1$; bottom row), moderately improves the true negative rate, but more importantly dramatically hampers the ability to detect non-null interactions for small interaction strengths. A value of $\gamma = 0.5$ seems to be a good compromise allowing to efficiently detect true interactions when their strength is not too small (at least 0.1 day^{-1}) but avoiding the identification of false interactions. It gives good results especially when

311 the sampling lasts more than 400 trapping days per camera. Hence, we decided to use a penalization
312 parameter of $\gamma = 0.5$ to infer the parameters of a MHP from real data (see Section 3.2). Lastly,
313 supplementary analyses show that the spurious interactions are randomly distributed and not biased towards
314 indirect interactions as with the inter-event times method (Appendix S1: [Section S1](#)).

315 **3.2 Analysis of real data**

316 We fitted a MHP using the occurrence data of five species (impala, greater kudu, lion, blue wildebeest and
317 Burchell's zebra) collected with camera traps. Adjusting the model only took a few seconds on a personal
318 computer. The resulting interaction functions are shown in Figure 5 and the inferred background rates in
319 Appendix S1: [Section S2](#).

320 Background rates represent the basal intensity for each species, independently of the others. They vary
321 greatly between species, with impala having a much higher background rate than other species (impala:
322 0.212 day^{-1} ; zebra: 0.040 day^{-1} ; kudu: 0.035 day^{-1} ; wildebeest: 0.022 day^{-1} and lion: 0.003 day^{-1}). As
323 expected, they are strongly related to the total occurrence count of each species.

324 Regarding the interaction functions, the inferred parameters highlight a strong auto-attraction for the first
325 bin (0-6h), varying between 1.5 and 2.25 day^{-1} depending on the species. Regarding the cross-species
326 interaction functions, many herbivores are attracted to each other. Impalas follow or avoid kudus
327 (depending on the delay), wildebeests and zebras; zebras follow impalas, kudus and wildebeests;
328 wildebeests mainly follow zebras. Other interactions between herbivores are negligible. These
329 herbivore-herbivore interactions are composed of a short-term attraction (during the first six hours after an
330 occurrence) and of a medium-term attraction (twelve to thirty-six hours after an occurrence), except
331 impalas that are not attracted by zebras on the short-term. Additionally, impalas seem to avoid kudus six to
332 twelve and thirty to thirty-six hours after an occurrence. We notice that these interactions are asymmetrical
333 (impalas and zebras follow other species much more than they are followed). Regarding prey-predator
334 interactions, lions do not follow or avoid any other species. Zebra and impala seem to avoid lion in the next

335 6 hours following an occurrence of this predator. Finally, the inferred interactions are relatively robust to a
336 change in bin width, as we show in Appendix S1: [Section S4](#), where we performed the inference on the
337 same dataset with different bins widths (3 and 9 hours).

338 **4 Discussion**

339 It is now well established that identifying the signature of inter-specific interactions from species
340 occurrence data is generally difficult ([Blanchet et al., 2020](#); [Popovic et al., 2019](#)). However, camera trap
341 data provide additional information (time and order of occurrence of species) that can help to relate
342 occurrence patterns to underlying inter-specific interactions. In this context, the Hawkes model provides a
343 new theoretical framework to analyze species occurrences sampled in continuous time using camera traps.
344 This model aims to predict the probability of occurrence of a given species at a given time taking into
345 account the previous occurrences for several species. By considering the exact time at which species occur,
346 this model provides a detailed picture of species reactive temporal interactions under the form of interaction
347 functions (here, the term “interaction” refers to the attraction/repulsion pattern). These functions allow a
348 multiscale description of interactions as they characterize how the interaction strength varies with time,
349 contrary to other methods that provide a single measure of attraction/avoidance. Moreover, these functions
350 are directed: the inferred interactions can be asymmetrical, as expected for ecological interactions. The
351 toolbox associated to this model offers the possibility to generate data and design simulation studies or to
352 infer parameters from real data.

353 We used the multivariate Hawkes process to generate camera trap-like datasets with different properties
354 (sampling effort, strength of interactions) and showed how these simulated data can be used to evaluate the
355 performance of a method or to tune inference parameters. Both simulation studies demonstrate that camera
356 trap data can be used to detect reactive temporal interactions between species but this requires a substantial
357 sampling effort, especially when the strengths of interactions are low. In our simulation setup (five species,
358 background rates of 0.1 day^{-1} and only attractions), results suggest that at least one year of sampling with

359 25 cameras is required to obtain reliable inference, and this holds only if the interaction is strong enough
360 (interactions of strength 0.01 day^{-1} are not reliably detected in our simulations). These requirements would
361 probably be higher if more species were considered, especially for rare species (smaller background rate).
362 Hence, we agree with [Schliep et al. \(2018\)](#) that more data are needed to estimate reliable reactive
363 interaction patterns than to estimate species occupancy. In this context, the MHP provides a powerful
364 simulation tool to design and assess the quality of sampling protocols in camera trap studies by adopting a
365 virtual ecologist approach ([Zurell et al., 2010](#)). The simulation study also highlights the limits of methods
366 focusing on pairs of species to analyze interactions between multiple species. By focusing only on two
367 species at a time, these approaches are not able to disentangle direct interactions from indirect effects due to
368 other species in interaction chains (Appendix S1: [Section S1](#)). Moreover, the correction for multiple testing
369 we applied in our study was not sufficient to eliminate these spurious interactions, and we can assume that
370 this issue is more important in the literature when no correction is considered. As a consequence,
371 inter-event times methods tend to overestimate the number of interactions, especially when their strength is
372 high or the sampling effort increases. However, such spurious interactions were inferred only when we
373 simulated quite strong interactions, and more investigations would be needed to estimate the range of
374 interaction strengths we can expect in natural conditions. On the contrary, the Hawkes process used here is
375 multivariate by nature, so it works on all species simultaneously and thus allows to identify interactions
376 between two species conditionally to the other species, similar to graphical models in the context of
377 co-occurrence analysis ([Popovic et al., 2019](#)). This modeling approach thus provides a better picture of the
378 interaction network of the whole community.

379 The real dataset analysis shows how the MHP can be used to infer reactive interactions between five
380 mammal species from the African savanna. In our example, since we defined an occurrence as the presence
381 of an individual or a group of individuals, the values of the interaction functions represent the number of
382 individuals or groups of individuals that are attracted/repulsed by other occurrences, and the typical group
383 size to consider depends of the species. We identified strong auto-attractions for all species but also

384 attractions between different herbivores and avoidance of lion by two herbivore species (impala and zebra).
385 Whereas it could be tempting to interpret these results as behavioral responses of species to an underlying
386 interaction (e.g. avoidance in response to predation), the Hawkes model only characterizes
387 attraction/avoidance patterns and particular care should be taken when interpreting these results, especially
388 since no covariates were included in this analysis. We discuss these different interpretations of the observed
389 patterns in terms of ecological processes below and we make suggestions to improve the MHP to untangle
390 the different hypotheses.

391 We identified auto-attractions for all species, indicating that the occurrence of a given species increases the
392 probability to have another occurrence of the same species at the same place. This could be due to the same
393 individual lingering in front of the camera, especially since no independence filter was applied (although
394 cameras are configured to pause for one minute between trigger events), or this could reflect sociality
395 among individuals, as an individual or a group may attract other individuals for gregarious species. This
396 could also stem from habitat selection processes, so that numerous subsequent occurrences could be
397 observed at cameras located in species' preferred habitats. Lastly, circadian rhythms impose physiological
398 constraints on the activity times of each species and thus could increase their occurrence rate at certain
399 times of the day. When they are not taken into account, as is the case here, circadian rhythms could affect
400 the interaction functions in the short term (0-6h) and also induce a 24-hour periodicity in the interaction
401 functions. This issue is clearly illustrated using simulated data (see Appendix S1: [Section S3](#)) and could
402 partly explain the short-term (0-6h) auto-attraction, and probably most of the weak auto-attraction observed
403 around 24h for impala, kudu, wildebeest and zebra (Figure 5).

404 Regarding the cross-species interactions, we observed attraction patterns between some herbivores, which
405 could be explained by four mechanisms. First, temporal niche convergence could induce attraction between
406 species when they are active at the same time of day and if they also share the same location. In our
407 example, the four herbivore species are diurnal with crepuscular activity peaks. However, if the apparent
408 attraction was due to shared circadian rhythms, we would probably observe a symmetry of interaction

409 functions (i.e., if $s_1 \rightarrow s_2$ is not null, $s_2 \rightarrow s_1$ is also not null) as the order of appearance of species at a
410 camera during the activity time would be random. This is not always the case in the example depicted here,
411 for instance between zebra and kudu. Second, species sharing the same kind of preferred environment
412 might show apparent attraction. However, as for the temporal niche, this spatial niche should induce a
413 symmetrical interaction pattern. Third, the apparent attraction between species could be due to
414 mixed-species grouping strategy, whereby some species forage together in mixed groups. Such groups are
415 thought to mitigate predation risk and/or to improve access to resources (Beaudrot et al., 2020). Moreover,
416 when it comes to predation risk, in addition to the dilution effect (a simple number game), there is a possible
417 benefit to being associated with more vulnerable species (Fitzgibbon, 1990). This implies a directionality
418 in the choices of association, an asymmetry well captured by the MHP. In our analysis, some species are
419 attracted by others in the first hours following an occurrence (impala follows kudu and wildebeest; zebra
420 follows wildebeest, kudu and impala; and wildebeest follows zebra). Interestingly, these associations have
421 been described in the literature (Meise et al., 2019; Pays et al., 2014; Schmitt et al., 2014). Finally, another
422 mechanism that could explain interactions between herbivore species is grazing succession (Bell, 1971),
423 which describes a strategy by which species sequentially use the same grazing area: less selective species
424 come first (non-ruminants and species with higher body mass), followed by more selective species (smaller
425 ruminants). In our results, some herbivore species are attracted with a delay (impala following zebra, kudu
426 and wildebeest; wildebeest following zebra; zebra following wildebeest and impala). Impala following
427 other (bigger) species and wildebeest following the non-ruminant zebra are compatible with the grazing
428 succession theory (Bell, 1971). However, the temporal scale of this potential grazing succession occurs at a
429 temporal scale much shorter than the one classically described (McNaughton, 1976, 1985).

430 Regarding the apparent avoidance of lions by zebras and impalas, here again this could stem from temporal
431 niche divergence (lion is a nocturnal species whereas impala and zebra are diurnal). This apparent
432 repulsion could also reflect a strategy of impala and zebra to minimize predation risk by reactively avoiding
433 lions, i.e. responding to actual cues of lions presence (olfactory or auditory cues for instance) at a fine

434 spatio-temporal scale, as documented for zebras ([Courbin et al., 2016](#)).

435 As discussed with the real dataset analysis, a major challenge remains linking attraction/repulsion patterns
436 identified by the MHP to underlying ecological processes. To date, the implementation used in this paper
437 cannot include covariates to model variations in species' background occurrence rates. This calls for two
438 major improvements: first, we could include temporal covariates to account for the variation of species
439 occurrence rate through the day according to their diel cycle. Second, we could include environmental
440 covariates to account for species habitat preferences across the landscape. Works such as [Fujita et al.](#)
441 [\(2018\)](#) for temporal covariates or [Carstensen et al. \(2010\)](#) for temporal and environmental covariates could
442 be helpful in this perspective. Further developments include accounting for the imperfect detection by
443 camera traps, which is known to be an important issue ([Burton et al., 2015](#)). In this regard, [Kellner et al.](#)
444 [\(2022\)](#) recently developed an occupancy model with a detection process occurring in continuous time with
445 a Markov-modulated Poisson process, and a similar approach could be envisioned with the MHP.

446 Here, we inferred a MHP from camera trap data, but this modeling approach could be extended to other
447 types of passive sensors collecting occurrence data in continuous time (e.g., microphones, hydrophones)
448 that are increasingly used to monitor biodiversity. In particular, using a spatially explicit extension of the
449 Hawkes process (first described by [Ogata, 1998](#), in the context of earthquakes occurrences) could be
450 especially suited to include a spatial dependency between camera traps or to analyze GPS collar data and
451 estimate interaction functions in time and space.

452 The Hawkes process could also be used for other applications than estimating inter-specific interactions, for
453 instance to study behavioral synchrony within a group (e.g. [Pays et al., 2012](#)) or to infer animal social
454 networks from occurrence data (e.g. [Jacoby et al., 2016](#)).

455 Even if more developments are required to improve ecological inference, we contend that the MHP and
456 other point processes methods offer an adapted theoretical framework for the analysis of time-continuous
457 occurrence data while contributing to an explanation of interactions among herbivores and between
458 herbivores and predators.

5 Acknowledgements

We thank all the people involved in Snapshot Safari data collection and management, including the students and volunteer groups who contribute to maintain the camera trap grids and the reserve owners and managers for opening their reserves. We also thank all Zooniverse volunteers who contributed to classify pictures. We thank Gilles Scarella and Patricia Reynaud-Bouret, developers of the `UnitEvents` package, for their detailed answers to our questions regarding the use of their package. We thank Mahendra Mariadassou, Vincent Miele, Sara Puijalon and Stéphane Robin for their input on this work during discussions. We thank Stéphane Delmotte and Adil El Filali for their input on the containerization of the code. We are also grateful to four anonymous reviewers whose comments substantially improved the manuscript. This work was partially performed using the computing facilities of the CC LBBE/PRABI. This work was partially supported by the grant ANR-18-CE02-0010 of the French National Research Agency ANR (project EcoNet).

6 Conflict of interests

The authors declare no conflict of interest.

References

- Albert, M., Y. Bouret, J. Chevallier, M. Fromont, F. Grammont, T. Laloe, C. Mascart, P. Reynaud-Bouret, A. Rouis, G. Scarella, and C. Tuleau-Malot, 2021. `UnitEvents`: Unitary Events Method with Delayed Coincidence Count (MTGAUE or Permutation Method) and Bernstein Lasso method for Hawkes processes. https://sourcesup.renater.fr/frs/?group_id=3267.
- Beaudrot, L., M. S. Palmer, T. M. Anderson, and C. Packer. 2020. Mixed-species groups of Serengeti grazers: a test of the stress gradient hypothesis. *Ecology* **101**:e03163.
- Bell, R. H. 1971. A grazing ecosystem in the Serengeti. *Scientific American* **225**:86–93.

481 Blanchet, F. G., K. Cazelles, and D. Gravel. 2020. Co-occurrence is not evidence of ecological
482 interactions. *Ecology Letters* **23**:1050–1063.

483 Burton, A. C., E. Neilson, D. Moreira, A. Ladle, R. Steenweg, J. T. Fisher, E. Bayne, and S. Boutin. 2015.
484 Wildlife camera trapping: A review and recommendations for linking surveys to ecological processes.
485 *Journal of Applied Ecology* **52**:675–685.

486 Caravaggi, A., P. B. Banks, A. C. Burton, C. M. V. Finlay, P. M. Haswell, M. W. Hayward, M. J. Rowcliffe,
487 and M. D. Wood. 2017. A review of camera trapping for conservation behaviour research. *Remote*
488 *Sensing in Ecology and Conservation* **3**:109–122.

489 Carstensen, L., A. Sandelin, O. Winther, and N. R. Hansen. 2010. Multivariate Hawkes process models of
490 the occurrence of regulatory elements. *BMC Bioinformatics* **11**:456.

491 Cornhill, K. L., C. Kelly, and G. I. H. Kerley. 2022. Lion reintroduction demonstrates that resident cheetah
492 have a spatially reactive response to lion. *African Journal of Ecology* **60**:1–12.

493 Cornhill, K. L., and G. I. H. Kerley. 2020. Cheetah communication at scent-marking sites can be inhibited
494 or delayed by predators. *Behavioral Ecology and Sociobiology* **74**:21.

495 Courbin, N., A. J. Loveridge, D. W. Macdonald, H. Fritz, M. Valeix, E. T. Makuwe, and
496 S. Chamaillé-Jammes. 2016. Reactive responses of zebras to lion encounters shape their predator-prey
497 space game at large scale. *Oikos* **125**:829–838.

498 Cusack, J. J., A. J. Dickman, M. Kalyahe, J. M. Rowcliffe, C. Carbone, D. W. MacDonald, and T. Coulson.
499 2017. Revealing kleptoparasitic and predatory tendencies in an African mammal community using
500 camera traps: a comparison of spatiotemporal approaches. *Oikos* **126**:812–822.

501 Daley, D. J., and D. Vere-Jones. 2003. *An introduction to the Theory of Point Processes*. 2nd edition.
502 Springer, New York.

503 Fitzgibbon, C. D. 1990. Mixed-species grouping in Thomson's and Grant's gazelles: the antipredator
504 benefits. *Animal Behaviour* **39**:1116–1126.

505 Frey, S., J. T. Fisher, A. C. Burton, and J. P. Volpe. 2017. Investigating animal activity patterns and

506 temporal niche partitioning using camera-trap data: challenges and opportunities. *Remote Sensing in*
507 *Ecology and Conservation* **3**:123–132.

508 Fujita, K., A. Medvedev, S. Koyama, R. Lambiotte, and S. Shinomoto. 2018. Identifying exogenous and
509 endogenous activity in social media. *Physical Review E* **98**:052304.

510 Galindo-Aguilar, R. E., B. C. Luna-Olivera, M. Ramírez-Ibáñez, and M. C. Lavariega. 2022.
511 Spatiotemporal co-occurrence of predators and prey in a neotropical mammal community in southern
512 Mexico. *Journal of Tropical Ecology* **38**:285–294.

513 Hansen, N. R., P. Reynaud-Bouret, and V. Rivoirard. 2015. Lasso and probabilistic inequalities for
514 multivariate point processes. *Bernoulli* **21**:83–143.

515 Harmsen, B. J., R. J. Foster, S. C. Silver, L. E. T. Ostro, and C. P. Doncaster. 2009. Spatial and Temporal
516 Interactions of Sympatric Jaguars (*Panthera onca*) and Pumas (*Puma concolor*) in a Neotropical Forest.
517 *Journal of Mammalogy* **90**:612–620.

518 Hawkes, A. G. 1971. Spectra of some self-exciting and mutually exciting point processes. *Biometrika*
519 **58**:83–90.

520 Hawkes, A. G. 2018. Hawkes processes and their applications to finance: a review. *Quantitative Finance*
521 **18**:193–198.

522 Hettena, A. M., N. Munoz, and D. T. Blumstein. 2014. Prey Responses to Predator’s Sounds: A Review
523 and Empirical Study. *Ethology* **120**:427–452.

524 Jacoby, D. M. P., Y. P. Papastamatiou, and R. Freeman. 2016. Inferring animal social networks and
525 leadership: applications for passive monitoring arrays. *Journal of The Royal Society Interface*
526 **13**:20160676.

527 Karanth, K. U., A. Srivathsa, D. Vasudev, M. Puri, R. Parameshwaran, and N. S. Kumar. 2017.
528 Spatio-temporal interactions facilitate large carnivore sympatry across a resource gradient. *Proceedings*
529 *of the Royal Society B: Biological Sciences* **284**:20161860.

530 Kellner, K. F., A. W. Parsons, R. Kays, J. J. Millspaugh, and C. T. Rota. 2022. A Two-Species Occupancy

531 Model with a Continuous-Time Detection Process Reveals Spatial and Temporal Interactions. *Journal of*
532 *Agricultural, Biological and Environmental Statistics* **27**:321–338.

533 Kuijper, D. P. J., M. Verwijmeren, M. Churski, A. Zbyryt, K. Schmidt, B. Jędrzejewska, and C. Smit. 2014.
534 What Cues Do Ungulates Use to Assess Predation Risk in Dense Temperate Forests? *PLoS ONE*
535 **9**:e84607.

536 Lambert, R. C., C. Tuleau-Malot, T. Bessaih, V. Rivoirard, Y. Bouret, N. Leresche, and P. Reynaud-Bouret.
537 2018. Reconstructing the functional connectivity of multiple spike trains using Hawkes models. *Journal*
538 *of Neuroscience Methods* **297**:9–21.

539 McNaughton, S. J. 1976. Serengeti Migratory Wildebeest: Facilitation of Energy Flow by Grazing.
540 *Science* **191**:92–94.

541 McNaughton, S. J. 1985. Ecology of a Grazing Ecosystem: The Serengeti. *Ecological Monographs*
542 **55**:259–294.

543 Meise, K., D. W. Franks, and J. Bro-Jørgensen. 2019. Using social network analysis of mixed-species
544 groups in African savannah herbivores to assess how community structure responds to environmental
545 change. *Philosophical Transactions of the Royal Society B* **374**:20190009.

546 Mohler, G., J. Carter, and R. Raje. 2018. Improving social harm indices with a modulated Hawkes process.
547 *International Journal of Forecasting* **34**:431–439.

548 Murphy, A., D. R. Diefenbach, M. Ternent, M. Lovallo, and D. Miller. 2021. Threading the needle: How
549 humans influence predator–prey spatiotemporal interactions in a multiple-predator system. *Journal of*
550 *Animal Ecology* **90**:2377–2390.

551 Nicvert, L., S. Donnet, M. Keith, M. Peel, M. J. Somers, L. H. Swanepoel, and J. A. Venter, 2023. Code
552 and data for: Using the multivariate Hawkes process to study interactions between multiple species from
553 camera trap data. <https://doi.org/10.6084/m9.figshare.24552157.v3>.

554 Niedballa, J., A. Wilting, R. Sollmann, H. Hofer, and A. Courtiol. 2019. Assessing analytical methods for
555 detecting spatiotemporal interactions between species from camera trapping data. *Remote Sensing in*

556 Ecology and Conservation **5**:272–285.

557 Ogata, Y. 1988. Statistical Models for Earthquake Occurrences and Residual Analysis for Point Processes.
558 Journal of the American Statistical Association **83**:9–27.

559 Ogata, Y. 1998. Space-Time Point-Process Models for Earthquake Occurrences. Annals of the Institute of
560 Statistical Mathematics **50**:379–402.

561 Ovaskainen, O., G. Tikhonov, A. Norberg, F. Guillaume Blanchet, L. Duan, D. Dunson, T. Roslin, and
562 N. Abrego. 2017. How to make more out of community data? A conceptual framework and its
563 implementation as models and software. Ecology Letters **20**:561–576.

564 Palmer, M. S., K. M. Gaynor, J. A. Becker, J. O. Abraham, M. A. Mumma, and R. M. Pringle. 2022.
565 Dynamic landscapes of fear: understanding spatiotemporal risk. Trends in Ecology & Evolution
566 **37**:911–925.

567 Pardo, L. E., S. P. Bombaci, S. Huebner, M. J. Somers, H. Fritz, C. Downs, A. Guthmann, R. S. Hetem,
568 M. Keith, A. le Roux, N. Mgqatsa, C. Packer, M. S. Palmer, D. M. Parker, M. Peel, R. Slotow, W. M.
569 Strauss, L. Swanepoel, C. Tambling, N. Tsie, M. Vermeulen, M. Willi, D. S. Jachowski, and J. A. Venter.
570 2021. Snapshot Safari: a large-scale collaborative to monitor Africa’s remarkable biodiversity. South
571 African Journal of Science **117**:1–4.

572 Parsons, A. W., C. Bland, T. Forrester, M. C. Baker-Whatton, S. G. Schuttler, W. J. McShea, R. Costello,
573 and R. Kays. 2016. The ecological impact of humans and dogs on wildlife in protected areas in eastern
574 North America. Biological Conservation **203**:75–88.

575 Parsons, A. W., K. F. Kellner, C. T. Rota, S. G. Schuttler, J. J. Millspaugh, and R. W. Kays. 2022. The effect
576 of urbanization on spatiotemporal interactions between gray foxes and coyotes. Ecosphere **13**:e3993.

577 Pays, O., A. Ekori, and H. Fritz. 2014. On the Advantages of Mixed-Species Groups: Impalas Adjust Their
578 Vigilance When Associated With Larger Prey Herbivores. Ethology **120**:1207–1216.

579 Pays, O., E. Sirot, and H. Fritz. 2012. Collective Vigilance in the Greater Kudu: Towards a Better
580 Understanding of Synchronization Patterns. Ethology **118**:1–9.

581 Popovic, G. C., D. I. Warton, F. J. Thomson, F. K. Hui, and A. T. Moles. 2019. Untangling direct species
582 associations from indirect mediator species effects with graphical models. *Methods in Ecology and*
583 *Evolution* **10**:1571–1583.

584 Prat-Guitart, M., D. P. Onorato, J. E. Hines, and M. K. Oli. 2020. Spatiotemporal pattern of interactions
585 between an apex predator and sympatric species. *Journal of Mammalogy* **101**:1279–1288.

586 R Core Team, 2023. R: A language and environment for statistical computing. R Foundation for Statistical
587 Computing. <https://www.R-project.org/>.

588 Reinhart, A. 2018. A Review of Self-Exciting Spatio-Temporal Point Processes and Their Applications.
589 *Statistical Science* **33**:299–318.

590 Reynaud-Bouret, P., V. Rivoirard, and C. Tuleau-Malot, 2013. Inference of functional connectivity in
591 Neurosciences via Hawkes processes. Pages 317–320 in 2013 IEEE Global Conference on Signal and
592 Information Processing. IEEE, Austin, TX, USA.

593 Ridout, M. S., and M. Linkie. 2009. Estimating overlap of daily activity patterns from camera trap data.
594 *Journal of Agricultural, Biological, and Environmental Statistics* **14**:322–337.

595 Rota, C. T., M. A. R. Ferreira, R. W. Kays, T. D. Forrester, E. L. Kalies, W. J. McShea, A. W. Parsons, and
596 J. J. Millsbaugh. 2016. A multispecies occupancy model for two or more interacting species. *Methods*
597 *in Ecology and Evolution* **7**:1164–1173.

598 Say-Sallaz, E., S. Chamaillé-Jammes, H. Fritz, and M. Valeix. 2019. Non-consumptive effects of predation
599 in large terrestrial mammals: Mapping our knowledge and revealing the tip of the iceberg. *Biological*
600 *Conservation* **235**:36–52.

601 Schliep, E. M., A. E. Gelfand, J. S. Clark, and R. Kays. 2018. Joint Temporal Point Pattern Models for
602 Proximate Species Occurrence in a Fixed Area Using Camera Trap Data. *Journal of Agricultural,*
603 *Biological and Environmental Statistics* **23**:334–357.

604 Schmitt, M. H., K. Stears, C. C. Wilmers, and A. M. Shrader. 2014. Determining the relative importance of
605 dilution and detection for zebra foraging in mixed-species herds. *Animal Behaviour* **96**:151–158.

606 Searle, C. E., J. B. Smit, J. J. Cusack, P. Strampelli, A. Grau, L. Mkuburo, D. W. Macdonald, A. J.
607 Loveridge, and A. J. Dickman. 2021. Temporal partitioning and spatiotemporal avoidance among large
608 carnivores in a human-impacted African landscape. *PLoS ONE* **16**:e0256876.

609 Valiente-Banuet, A., M. A. Aizen, J. M. Alcántara, J. Arroyo, A. Cocucci, M. Galetti, M. B. García,
610 D. García, J. M. Gómez, P. Jordano, R. Medel, L. Navarro, J. R. Obeso, R. Oviedo, N. Ramírez, P. J.
611 Rey, A. Traveset, M. Verdú, and R. Zamora. 2015. Beyond species loss: The extinction of ecological
612 interactions in a changing world. *Functional Ecology* **29**:299–307.

613 Zurell, D., U. Berger, J. S. Cabral, F. Jeltsch, C. N. Meynard, T. Münkemüller, N. Nehrbass, J. Pagel,
614 B. Reineking, B. Schröder, and V. Grimm. 2010. The virtual ecologist approach: simulating data and
615 observers. *Oikos* **119**:622–635.

7 Figure captions

Figure 1: Example of a realization of a multivariate Hawkes process. (a) shows the interaction network between five species (each arrow represents a non-null interaction function). In this example, all auto-interaction functions $f_{i \rightarrow i}$ are null. (b) shows the shape of the interaction functions ($K = 12$ time bins of width $\delta = 4$ hours) corresponding to arrows in the interaction network (a). (c) shows a realization of the Hawkes process with the interaction network and the interaction functions shown in (a) and (b). For this simulation, the background rate was set to 0.2 occurrences day^{-1} for all species. For each species i , the above panel shows the intensity λ_i and the bottom panel shows the species occurrences. Each time an attracting species occurs, the intensity for the attracted species peaks and then decreases as dictated by the interaction function shape.

Figure 2: Study sites. Six protected areas were surveyed with camera traps for this study: the Associated Private Nature Reserves (APN), Kruger National Park, Madikwe Game Reserve, Pilanesberg National Park, Somkhanda Game Reserve and Venetia Limpopo Nature Reserve. Data © OpenStreetMap contributors.

Figure 3: Evaluation of an inter-event times method (Murphy et al., 2021). Panels represent different interaction strengths (maximum value of the interaction function). The x-axis represents the sampling length and the y-axis represents the performance: true positive rate (full dots, continuous line) or true negative rate (circles, dashed line). Points indicate values for the 30 repetitions, lines joins the medians, and the colored area represents the 2.5th and 97.5th percentiles.

Figure 4: Performance of the inference with the multivariate Hawkes model. In columns, the interaction strength (maximum value of the interaction function). In rows, the different values of the penalization parameter γ . The x-axis represents the sampling length and the y-axis represents the performance (true positive rate or true negative rate). Lines, points and colors have the same meaning as for Figure 3.

Figure 5: Inference of interactions from real data using the multivariate Hawkes model. The top plot shows the auto-interaction functions (between occurrences of the same species). The bottom plot shows

641 cross-species interactions, where the intensity of species in rows is affected by species in columns. The
642 horizontal dashed line represents zero. Note that the y-axis scale is different between auto- and
643 cross-species interactions. Silhouette images from [PhyloPic](#) by Lukasiniho (wildebeest), Margot Michaud
644 (lion), Robert Hering (kudu), Zimices (zebra) and an unknown author (impala).

8 Figures

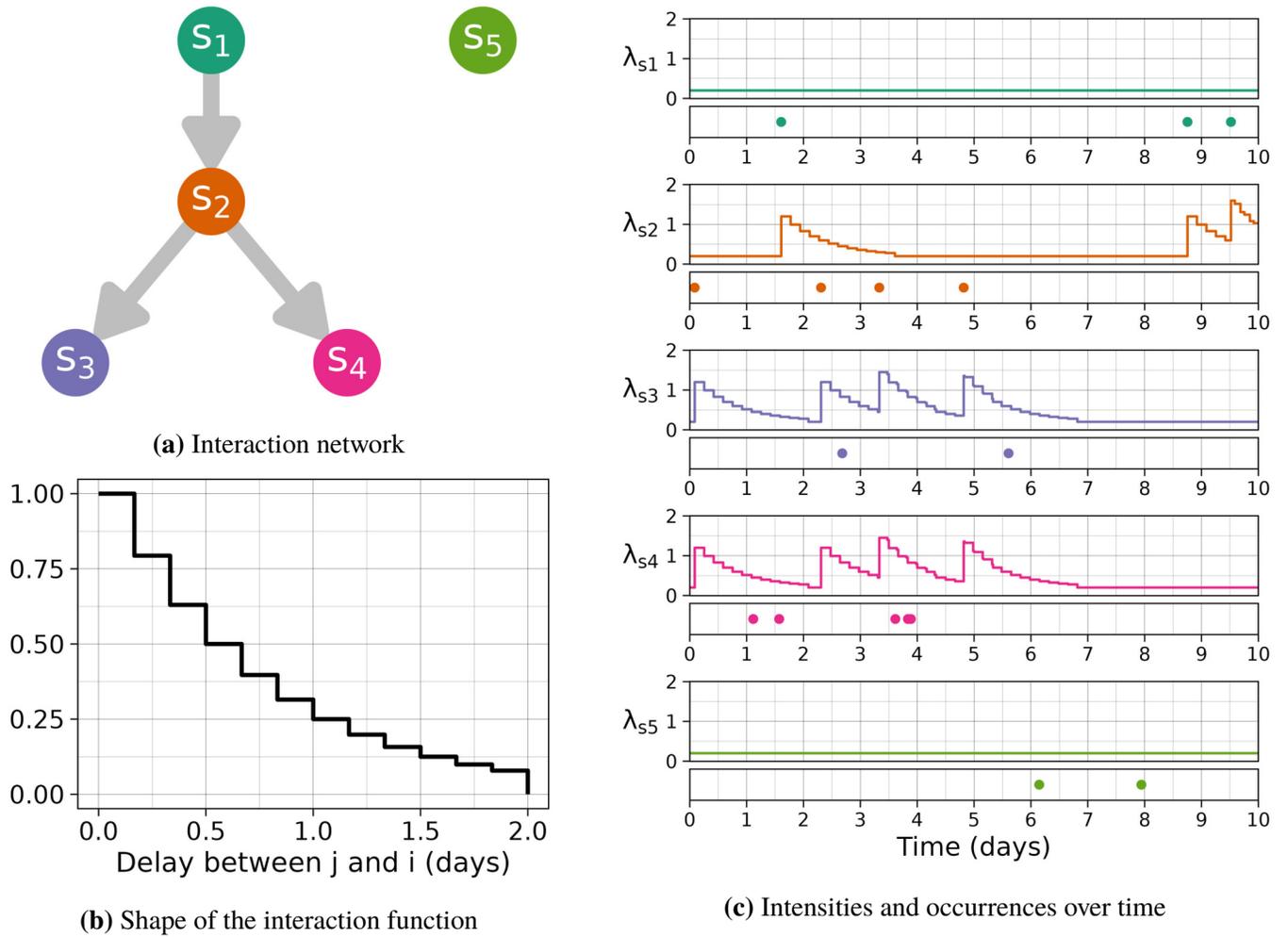


Figure 1

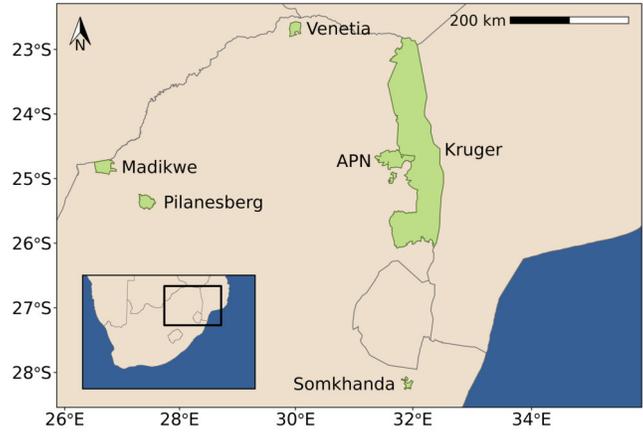


Figure 2

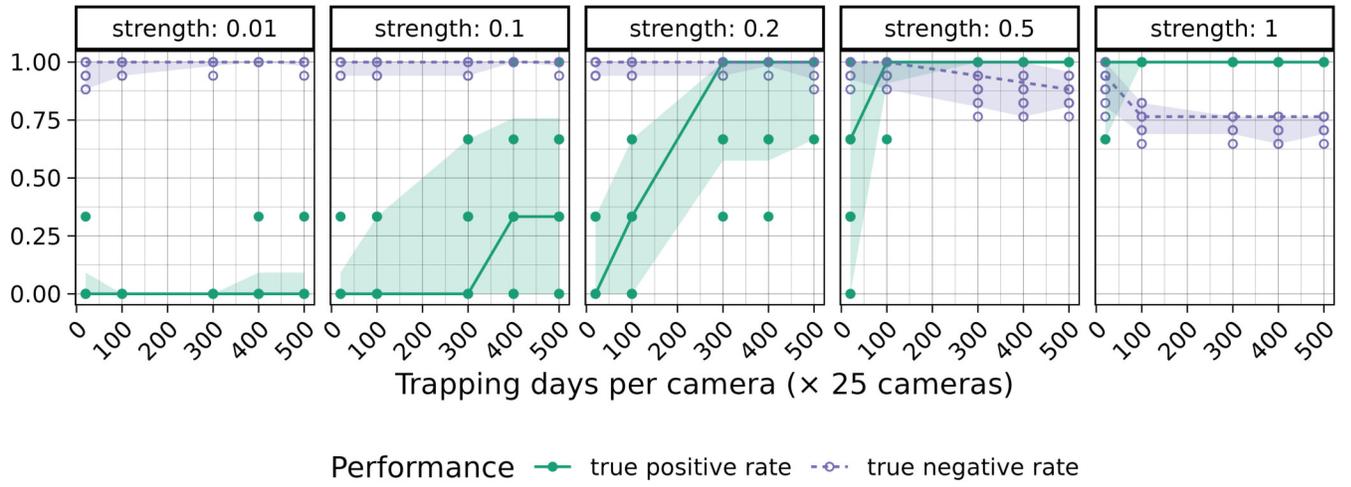


Figure 3

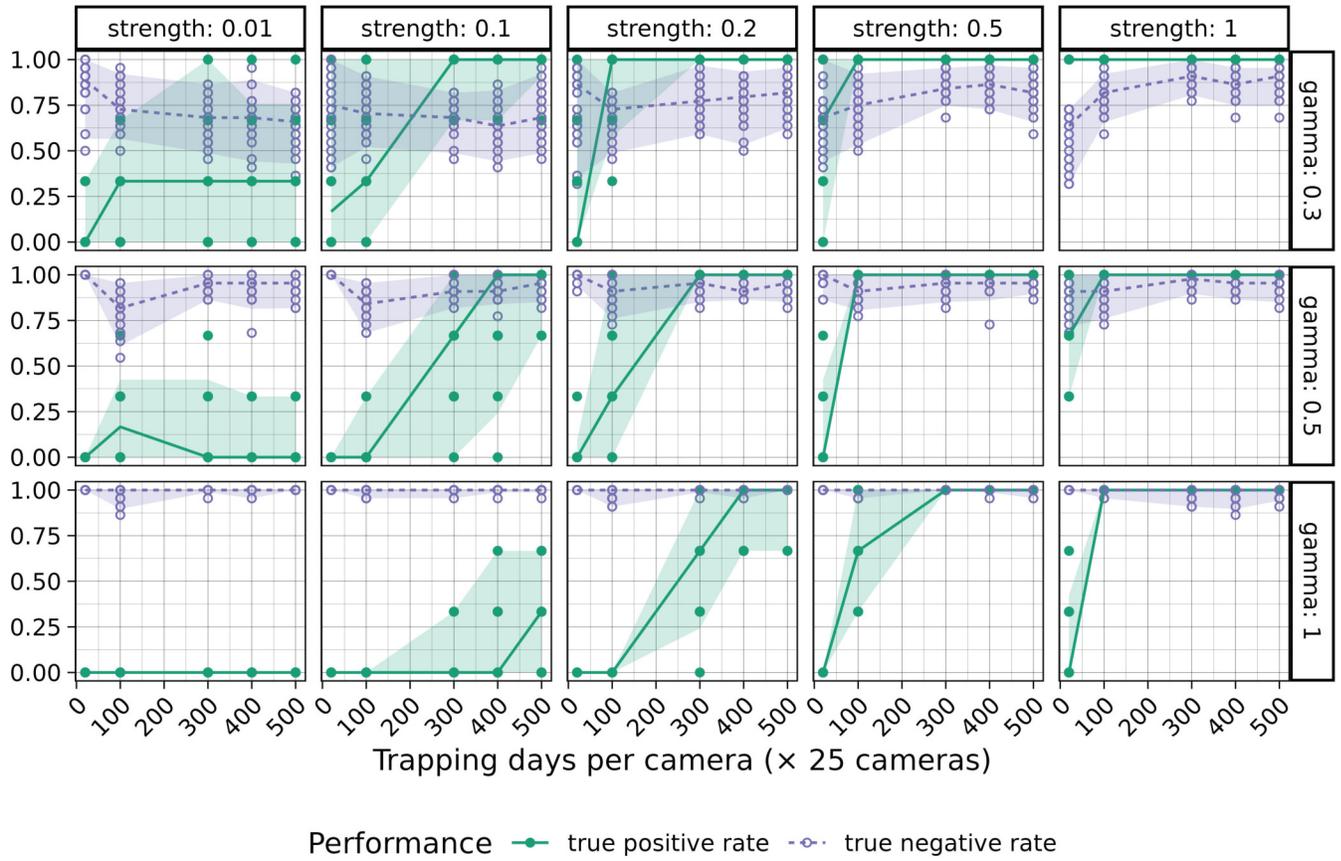
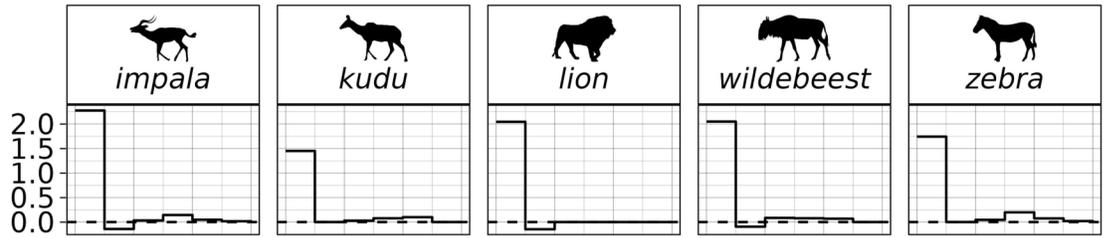


Figure 4

Auto-interactions



Cross-species interactions

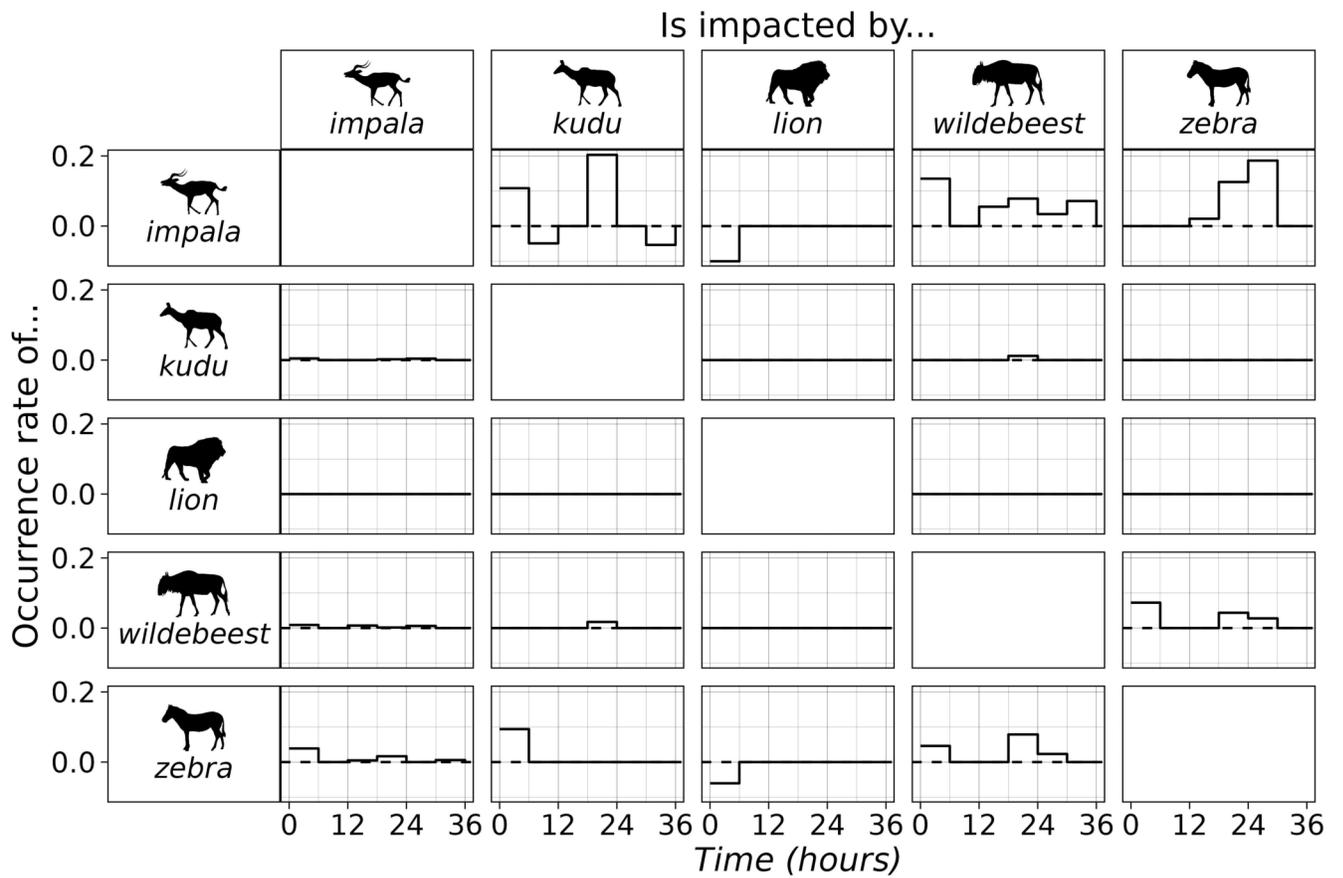


Figure 5

Ecology

Appendix S1

Using the multivariate Hawkes process to study interactions between multiple species from camera trap data

Lisa Nicvert¹, Sophie Donnet², Mark Keith³, Mike Peel^{4, 5, 6}, Michael J. Somers³, Lourens H. Swanepoel⁷, Jan Venter^{8, 9}, Hervé Fritz^{9, 10}, and Stéphane Dray¹

¹Université de Lyon, Université Lyon 1, CNRS, VetAgro Sup, Laboratoire de Biométrie et Biologie Evolutive, UMR5558, Villeurbanne, France

²Paris-Saclay University, AgroParisTech, INRAE, UMR MIA-Paris, France

³Eugène Marais Chair of Wildlife Management, Mammal Research Institute, Department of Zoology and Entomology, University of Pretoria, Pretoria, South Africa

⁴Agricultural Research Council, Animal Production Institute, Rangeland Ecology, Pretoria, South Africa

⁵School of Animal, Plant and Environmental Sciences, University of the Witwatersrand, Johannesburg, South Africa

⁶Applied Behavioural Ecology and Ecosystems Research Unit, University of South Africa, Florida, South Africa

⁷Department of Biological Sciences, Faculty of Science, Engineering and Agriculture, University of Venda, Thohoyandou, 0950, South Africa

⁸Department of Conservation Management, Faculty of Science, George Campus, Nelson Mandela University, George, South Africa

⁹REHABS, International Research Laboratory, CNRS-NMU-UCBL, Nelson Mandela University, George, South Africa

¹⁰Sustainability Research Unit, Nelson Mandela University, George, South Africa

Correspondence: Lisa Nicvert (lisa.nicvert@univ-lyon1.fr)

Authorship statement: Hervé Fritz and Stéphane Dray are co-last authors.

This PDF includes [Section S1](#) to [Section S4](#) from Appendix S1.

Section S1 Bias in the inference methods

We are interested in knowing whether spurious interactions are more likely to be inferred between species which are part of an interaction chain. To investigate this question, we analyze the inferences performed with the inter-event times method of [Murphy et al. \(2021\)](#) and with the Hawkes method (respectively sections 3.1.2 and 3.1.1 in the main text). The true interaction graph used for these simulations is depicted in Figure 1a in the main text (s_1 attracts s_2 , and s_2 attracts s_3 and s_4). The total number of repetitions over all conditions were respectively $N = 750$ for the inter-event times method (Figure S1a) (5 durations \times 5 strengths \times 30 datasets) and $N = 2250$ for the Hawkes model (Figure S1b) (5 durations \times 5 strengths \times 30 datasets \times 3 penalizations). With the inter-events times method, the model is biased towards inferring more often the indirect interactions $s_1 \rightarrow s_3$, $s_1 \rightarrow s_4$, $s_3 \rightarrow s_4$ and $s_4 \rightarrow s_3$. This is not the case with the Hawkes model.

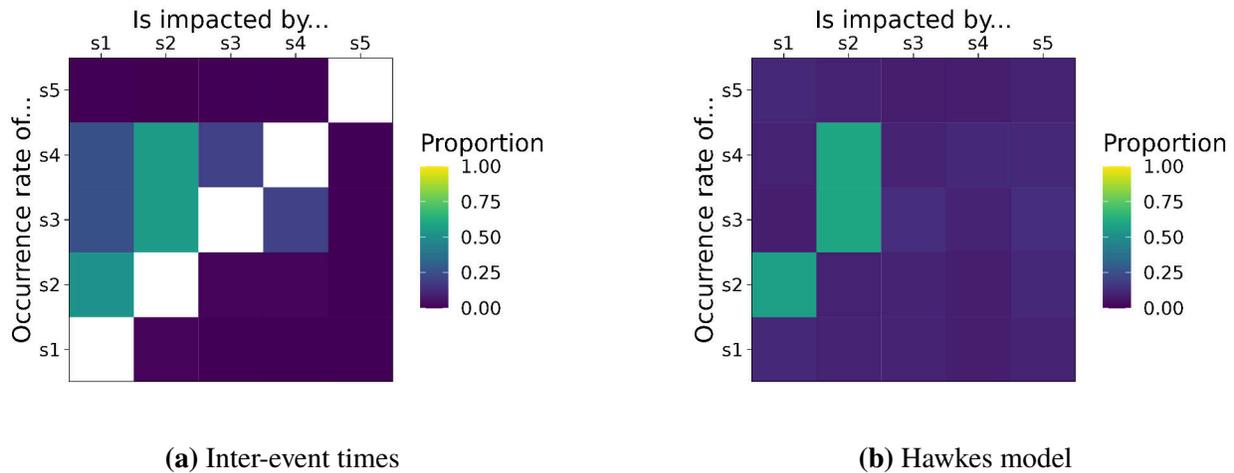


Figure S1: Bias in the inference. These graphs show the proportion of times an interaction was inferred over all simulated data in different conditions for (a) the inter-event times method of [Murphy et al. \(2021\)](#) and (b) the Hawkes model. The true interactions are $s_1 \rightarrow s_2$, $s_2 \rightarrow s_3$ and $s_2 \rightarrow s_4$.

Section S2 Background rates inferred for real data

The figure below shows the background rates inferred for each species in the application example. These background rates are closely linked to the species abundances observed on camera traps, with impala being by far the most abundant species (50 803 occurrences), followed by zebra (9 843), then wildebeest and kudu (respectively 5910 and 5560) and lion being far more rare (587).

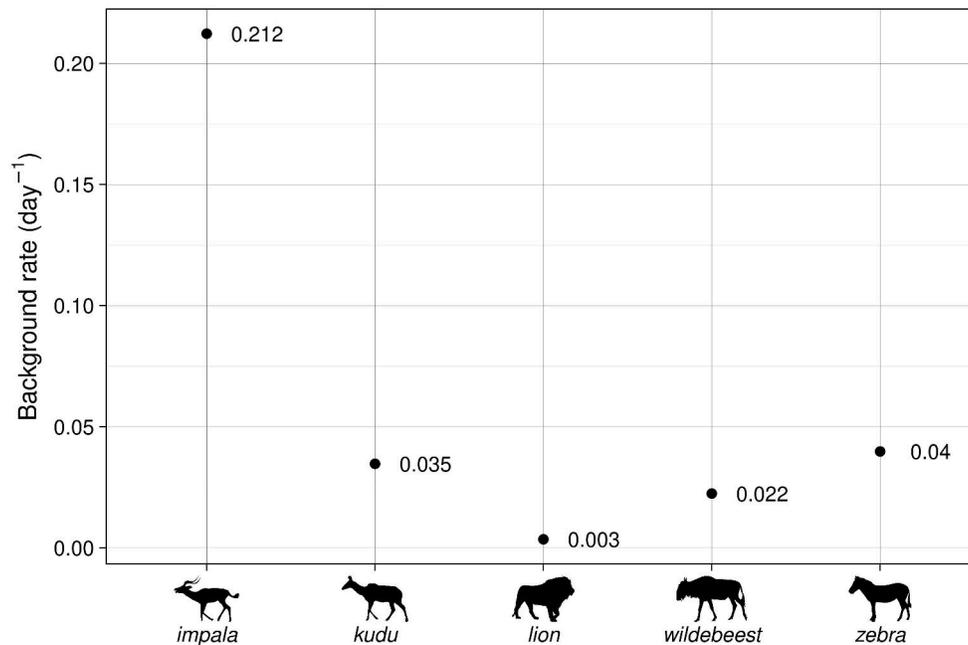


Figure S2: Background rates inferred with the Hawkes process for each species. The background rates values are written besides the points.

Silhouette images from [PhyloPic](#) by Lukasinio (wildebeest), Margot Michaud (lion), Robert Hering (kudu), Zimices (zebra) and an unknown author (impala).

Section S3 Influence of circadian rhythms

We simulated two species occurrences with cyclic rates in order to mimic circadian rhythms. We simulated occurrences over 1000 days on a single camera. For that, we used an inhomogeneous Poisson process for which species intensities varied periodically over 24 hours, with the Poisson non-homogeneous intensity defined with a cosine function with a period of 24 hours. Even though species occurrences were independent, the Hawkes process inferred a periodic attraction/repulsion pattern between and within species.

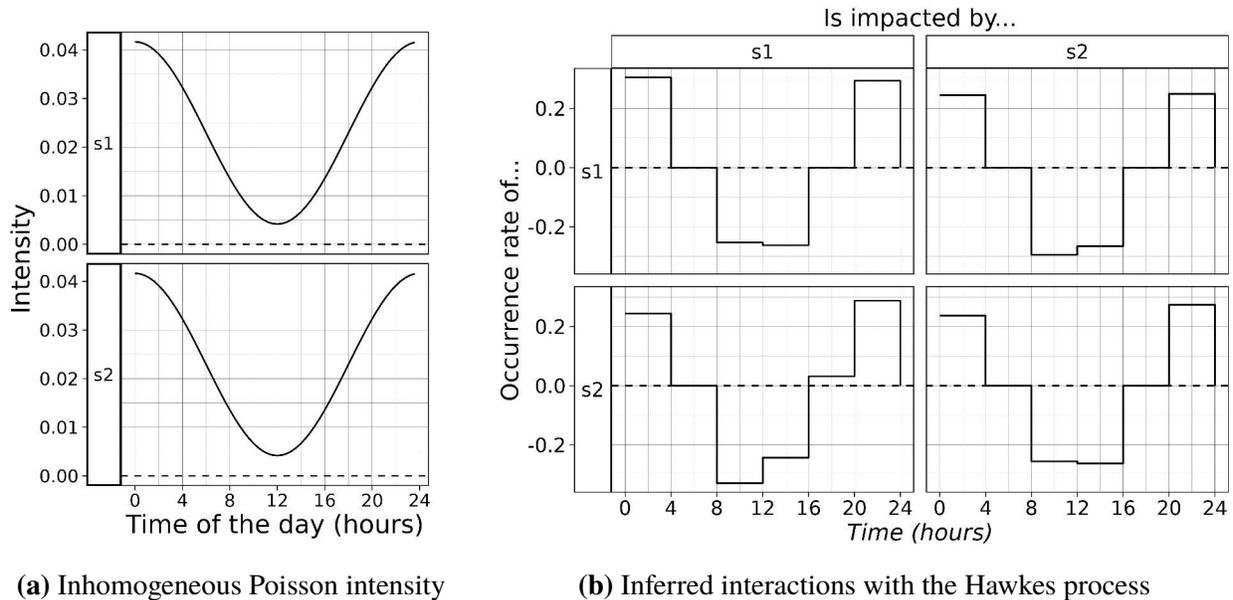


Figure S3: Inference of spurious interactions when circadian rhythms are present. (a) shows the non-homogeneous Poisson intensities for the two species. (b) shows the interaction functions inferred with a multivariate Hawkes process from the data simulated with two non-homogeneous Poisson process for which the intensities are shown in (a).

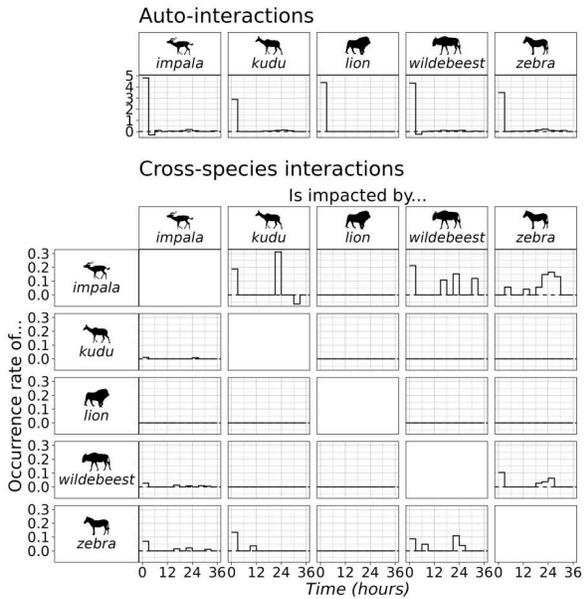
Section S4 Influence of bin width

In order to check if bin width influences the inferred interaction functions, we changed the bin width used for the inference on the real data. Here, we inferred a Hawkes process with bins widths δ of 3 and 9 hours (compared to $\delta = 6$ hours in the main text).

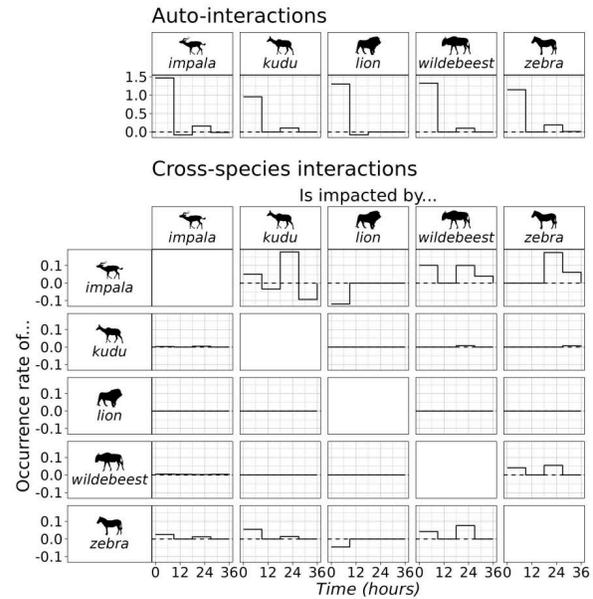
Results are shown in Figure S4. Overall, the parameters are similar between 3, 6 and 9 hour bins. Regarding the 3 hours bins, the nonzero interactions are the same than with 6 hours bins and they have the same direction. Only the negative interactions (lion-impala and lion-zebra) are missed, possibly due to low power. The inferred background rates are similar to the rates inferred with 6 hours bins, except for impala which has a smaller background rate (0.194 compared to 0.212 occurrences day^{-1} in the main text).

Regarding the 9 hours bins, all nonzero interactions are the same and they have the same direction as in the 6 hours bins version. The inferred background rates are similar to those inferred with 6 hours bins as well.

Finally, we can note that the values of the interaction functions are smaller when the bins widths are larger. This is expected because the average number of occurrences gained or suppressed during a given interval is the integral of the function over this interval.



(a) Bin width: 3 hours



(b) Bin width: 9 hours

Figure S4: Inference of a multivariate Hawkes process with different bin widths. (a) shows the parameters inferred with 3 hours bins and (b) the parameters inferred with 9 hours bins. Silhouette images from [PhyloPic](#) by Lukasinio (wildebeest), Margot Michaud (lion), Robert Hering (kudu), Zimices (zebra) and an unknown author (impala).

References

Murphy, A., D. R. Diefenbach, M. Ternent, M. Lovallo, and D. Miller. 2021. Threading the needle: How humans influence predator–prey spatiotemporal interactions in a multiple-predator system. *Journal of Animal Ecology* **90**:2377–2390.